- 1 Coupling beach ecology and macroplastics litter studies: current trends and the way ahead
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- 20 Abstract
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22 As sites of floating marine material deposition, sandy beaches accumulate marine litter. While 23 research and assessment on beach litter is increasing and involves various actors (scientists, society 24 and NGOs), there is the need to assess current and future dominant trends, directions and priorities 25 in that research. As such, a textural co-occurrence analysis was applied to published scientific 26 literature. Words were considered both singly and as part of compound terms related to concepts 27 relevant to sandy beach ecology: morphodynamic state; Littoral Active Zone; indicator fauna. Litter 28 as a compound term was also included. The main co-occurrences were found within compounds, 29 with scarce interaction of "morphodynamic state" with the others, indicating the need for further integration of beach ecology paradigms into beached plastics studies. Three approaches are 30 31 proposed to overcome the research limits highlighted: the unequivocation of terms, the 32 consideration of adequate scales, and the attention to dynamics rather than just patterns.

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34 Keywords

Beached macroplastics; co-occurrence analysis; Littoral Active Zone; Indicator fauna; Beach
 geomorphology

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38 1. Introduction

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40 It is widely accepted that marine litter is a global phenomenon, recognized of concern at 41 international levels therefore included in the UNEP initiatives such as the Sustainable Development 42 Goals (SDG) or in G7 and G20 statements (Borja and Elliott, 2019). Indeed, SDG14 (Life below 43 water) specifically has an extremely ambitious target to reduce or remove this source of pollution 44 by 2025 (UN, 2015) although without further development that target has been criticized as being 45 inaccurate and unattainable (Cormier and Elliott, 2017). Marine Litter has been defined by UNEP, 46 2005. as "any persistent, manufactured or processed solid material discarded, disposed of or 47 abandoned in the marine and coastal environment". Macroplastics are a component of plastic litter, 48 defined as plastic pieces above 25 mm size (Galgani et al., 2013), and further detailed as size-49 classes in the new guideline about macrolitter monitoring (Fleet et al., 2021). They include a broad 50 range of materials and shapes, due to production, mechanical alterations or differential weathering 51 and other degradation conditions of a complex of different polymers (Frigione et al., 2021). 52 Macroplastics litter is often the source of secondary microplastic contamination (Andrady, 2011, 53 Lambert et al., 2014, GESAMP, 2015). Although connected, research related to macroplastics litter 54 differs greatly from that of microplastics in terms of study design, protocols, and analyses (Fleet et 55 al., 2021). Addressing macroplastics contamination and pollution is likely to identify paths from 56 source of littering to the access to food webs via breakdown.

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58 1.1 Sandy beaches and beached plastics

59 Sandy beaches are an ecosystem exposed to and under threat from many global environmental 60 problems, notable those termed the *triple whammy* of increased urbanization and industrialization, 61 increased use of resources and decreased resistance and resilience to external threats such as climate 62 change (Defeo and Elliott 2021). The relatively young discipline (established in the 1980s, 63 McLachlan, 1983) of sandy shore ecology began by identifying features shaping those physically-64 driven environments, and then proceeded by overlapping morphodynamic characterisation with 65 biotic data layers, finally superimposed on by human pressures (Defeo et al., in press). Current paradigms define the morphodynamic type of a beach as the interaction between sand particle size 66 67 and exposure to tidal range and wave conditions: as such, dissipative beaches are characterized by 68 gentle slopes, wide beach width and fine grain sizes as relevant features. By contrast, the reflective 69 end of the scale occurs when the sediment is coarse and stored on the intertidal beach and 70 backshore, and where there is no surf zone and waves surge directly up the beach face (McLachlan 71 and Defeo 2018). The macrofauna inhabiting beach environments reflects these variations: an 72 increasing number of species is found toward dissipative beaches, which are more benign as less 73 exposed to substrate tumbling. With a progression through the morphodynamic spectrum through 74 intermediate beaches, most species become less successful, and few can colonize reflective beaches 75 due to the harsher environment given by the saltation of coarse substratum particles subjected to the 76 high energy of incoming waves. The morphodynamic state is hence relevant to beach functioning, 77 with direct repercussion on the quality and quantity of ecosystem services (McLachlan et al. 2013; 78 McLachlan and Defeo 2018). Consequently, the occurrence of beached plastic could also be 79 affected by the different exposure to and interaction with energy, matter and biota. The co-80 occurrence of environmental features and beached plastics data could reveal potential interactions 81 occurring on matching spatial and/or temporal scales. It is hence timely to propose tools and 82 standards quantifying beached plastic and beach ecological processes. For instance, the average 83 specific gravity of plastics and polymers is 1.275±0.303 g.cm⁻³ (calculated from AmesWeb, 2021) 84 whereas that of substratum particles such as quartz grains is 2.65 g.cm⁻³ and that of marine mollusc 85 shells 2.68-2.72 g.cm⁻³. Therefore, plastic and polymer accumulate, are buried and re-suspended 86 (Williams and Tudor, 2001). Density, shape and relative size of macroplastics and substratum 87 particles are important when considering these dynamics, occurring along the land-sea axis 88 (Lebreton et al. 2019; Rangel-Buitrago et al., 2017; Moreira et al., 2016; Cresta and Battisti, 2021). 89 Given the high relevance of the local level of beaches (Fanini et al., 2020), the variability in 90 substratum and exposure will likely require tailored approaches depending on morphophysical and 91 landscape features (Ryan and Perold., 2021) together with the application of standard protocols, 92 essential to achieve data interoperability.

93 Macroplastics is also the most common subject of beach clean-up activities or citizen observation-94 based initiatives and monitoring actions. There is a common top-down approach to the topic, 95 engaging society as citizen scientists and monitors (see the definition by ECSA, European Citizen 96 Science Association http://ecsa.citizen-science.net/:). NGOs, private sectors and national agencies 97 and departments are conducting surveys, campaigns and projects supporting data collections and 98 evidence-based policies (Hidalgo-Ruz and Thiel, 2015; GESAMP, 2019; Syberg et al, 2020). 99 Despite this, there are still challenges in the definition of the role of citizen science and data that it 100 can provide (Haklay et al. 2021). However, it is through these activities that relevant evidence can 101 be built, enabling macroscale patterns to be determined and finally be mainstreamed into policies. 102 Indeed, it is through these campaigns that plastics were identified as the most common material 103 composing human litter on the beach (Addamo et al. 2017). Also, bans on single use plastic items 104 (SUP) were generally based on the top items found as beached macroplastics litter, on data 105 collected by citizens and mediated by NGOs in their mainstream to policy making. Country 106 implementation of international strategies such as the Programme of Measures for the European 107 Marine Strategy Framework Directive (MSFD) - of which marine litter is Descriptor number 10 for 108 determining Good Environmental Status - are also based on volunteer-led data collection. For 109 example, the main marine litter monitoring in the UK has been the annual volunteer-led beach clean 110 of the Marine Conservation Society (MCSUK) involving many thousand volunteers every 111 September since 1994; this was recognized as part of the UK contribution to implementing the 112 MSFD.

113 While these studies are powerful in depicting patterns and they strongly support governance via 114 evidence-based information, studies tackling dynamics remain limited. Such studies would require 115 the consideration of marine litter across temporal scales and disciplines, but also would need to be 116 based on shared and quality-assured protocols, and shared data, which are a frequent constraint in large-scale studies (but see Morales-Caselles et al., 2021). The temporal dimension in particular 117 reveals notable gaps, especially related to long-term designs and/or before-after impacts such as 118 119 floods, typhoons, and bans of specific items e.g. single use plastic bags. Again, the relevance of the 120 single beach unit in both social and ecological perspective would require attention since the very 121 planning of actions.

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123 1.2 Assessing plastics on beaches: protocols and state of the art

124 Reviews about methodologies for marine litter monitoring started in the 1990s (Rees and Pond, 125 1995) and standard methodologies are proposed by the Regional Seas Convention guidelines within 126 their action plans such as Cheshire et al., 2009 (UNEP/IOC), Helsinki Convention (HELCOM, 127 2008), OSPAR Commission (2010) and Schulz et al. (2017). Furthermore, monitoring guidelines 128 have been outlined for programmes such as the MSFD (Galgani et al., 2013), to support marine 129 litter baselines (Hanke et al. 2019), threshold values (Van Loon et al., 2020) or providing 130 harmonized list of items (Fleet et al., 2021). They mainly address: 1) Quantification (database -131 number, weight or volume);

132 2) Characterization (composition - master lists); and 3) Evidence-based policies for production
133 consumption systems (e.g. brand auditing, target items campaigns, or littering sources).

Selection criteria for beaches to be monitored are also given, both in the framework of national programmes (Opfer et al., 2012), or international regulations such as MSFD (EC 2008 2008/56/EC), where marine litter represents an indicator of the environmental quality status of the ecosystem. As a general approach, a set of desirable characteristics is provided for identifying the
sampling area to design monitoring and assessment programs as well as for beach cleanup
initiatives with volunteers (OSPAR, 2010; Galgani et al, 2013; GESAMP, 2019; WIOMSA manual
-Western Indian Ocean Marine Science Association (Barnardo and Ribbink, 2020)), and
UNEP/IOC manual (Cheshire et al., 2009).

142 In order to create robust and comparable quality-assured data, the monitoring methods have to be 143 standardized, agreed and implemented consistently. When this relates to the areas that are 144 monitored, the general indications about site selection include: accessibility of the site, and 145 avoidance of steep slopes (15°-45°); areas not subjected to cleaning activities; avoid nesting sites 146 for threatened species or presence of endangered or protected species; avoid streams, and natural or 147 artificial elements likely to interfere with currents. In particular, the WIOMSA manual suggests a 148 random selection of sites, and if this is not possible, a site selection guided by a pre-defined 149 criterion, without previous investigation, in essence having a random sampling design. In all cases, 150 the surveys for marine plastic macrolitter standing stock should be carried out along a 151 predetermined length of 100 m running parallel to the shoreline (Barnardo and Ribbink, 2020).

152 There are a few protocols adapted to beach morphology, such as considering whether the area is 153 macro or microtidal, and has reflective or dissipative conditions; fine sand or coarse sand or 154 pebbles, presence/absence of organic litter (GESAMP, 2019). Turra et al. (2014) called for 155 protocols relevant to sandy beach ecology (see also Moreira et al., 2016). However, to date, the 156 integration of relevant sandy beach variables is left to single initiatives rather than embedded in 157 protocols. However, beach structural features are intrinsically connected to functional processes 158 occurring around sandy shores, from physical and biotic (faunal) conditions to socio-economic 159 dimensions (see McLachlan and Defeo, 2018 for a recent comprehensive summary). For this 160 reason, a greater connection of beach ecology with plastic studies would increase the relevance of 161 research and enhance the support to policy and citizens. Given the high attention on the topic and 162 the response by scientists which produce much literature about marine plastics litter, it was urgent 163 to detect and communicate trends for future research. On this background, and with explicit focus 164 on the macroplastics fraction, the aim here is to show the integration of ecological features of sandy 165 beach systems into beached plastics litter studies. As such, the analysis of word co-occurrence in 166 scientific publications was identified as suitable first step in this process.

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168 2. Materials and Methods

169 2.1. Keywords, compounds and co-occurrence in scientific literature

This analysis starts from the attention to concepts and related keywords, as this is the background for any further data organization and analysis. The textural approach of word co-occurrence analysis of published literature has proven to be insightful across scientific disciplines (Callon et al., 1983), including ecology (Neff and Corley, 2009). This approach was found relevant in identifying trends and gaps in research on different topics; here, it was applied to a range of keywords extracted from both beach ecology and plastic litter studies, as follows:

176 1) identification of keywords related to beach features relevant to geomorphology, ecology and 177 biota, and of keywords related to beached plastic sizes (including macroplastics); 2) bibliometric 178 analysis of how often, in published literature, these words co-occur; 3) recommendations on 179 strategies and parameters to be applied within projects related to beached macroplastics litter.

We emphasise that this approach integrates beach ecology standard terms and concepts into marine macroplastics litter studies, and vice-versa (in a range of actions from research to opportunistic gapfilling visits, to citizen science campaigns and governance support). This has the added benefit of proactively and concurrently making data interoperable and beneficial to science and society.

In the synthesis here, given that globally relevant beach features extracted from the ecology of sandy shores are non-independent from each other, we therefore defined components as entities composed of a set of non-independent parts. This established a dimensionality in the exploration process, in a hierarchy defined by single keywords and compounds to which the keywords belong. Compounds and selected words were within the following categories:

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190 Litter. The meaning of "litter" includes but is not limited to anthropogenic litter, which also is not 191 exclusively related to plastics (Rangel-Buitrago et al., 2017). In the context of our analysis, litter 192 was considered as a compound term, and given our intended focus on macroplastics litter, keywords 193 were selected depending on the range of sizes most commonly used and standardized within plastics 194 studies (Frias and Nash, 2019). Different plastic sizes are non-independent when considering 195 weathering and breakdown, which are likely to occur on a beach, thereby creating secondary 196 particles. Standing stock is a term originally related to biomass, but increasingly used to assess 197 beached litter. It is specifically referred to a one-off count of beach plastics litter, and mentioned as 198 such in international protocols and guidelines (JRC, 2013). It was therefore included in the 199 compound.

The compound term "Litter" included the following keywords: "Plastic"; "Macroplastic";
"Microplastic"; "Nanoplastic"; "Macrolitter"; "Microlitter"; "Standing stock".

203 Morphodynamic state. The morphodynamic state is defined by sand, waves and tides and these two 204 last are related to beach exposure; in turn, exposure relates to fetch distance and wind speed and 205 direction. This state directly influences the human use of beaches, both individually and collectively 206 through their determination of beach morphodynamic types (McLachlan et al. 2018). 207 Morphodynamic variables are non-independent from each other, and a subtle combination of them 208 categorizes each state, from dissipative to reflective. It can thus be hypothesized that on beaches, 209 marine litter deposition, breakdown, resuspension and washing are also physically driven, likely by 210 a set of physical variables largely overlapping to those characterizing morphodynamic states.

The compound term named "Morphodynamic" included the following keywords: "Beach
exposure"; "Beach width"; "Beach slope"; "Grain size".

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214 The Littoral Active Zone (LAZ). The LAZ concept was introduced as a budgetary approach to 215 substratum dynamics (Tinley, 1985). A LAZ is composed by zones characterised by the dynamic 216 exchange of mobile substratum, hence the LAZ is connecting the subtidal to the littoral and to the 217 primary dune (Figure 1.). Recent extensions of the concept brought attention, from an initial energy 218 and substratum consideration only, to the resident fauna behavior and to the social and ecological 219 components of the system (Scapini et al., 2019; Fanini et al., 2021 respectively; Defeo et al. in 220 press). The functionality of a beach is tied to the LAZ, and a functional LAZ is conferring resilience 221 on the system.

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Figure 1. Schematic representation of the Littoral Active Zone and keywords extracted for the analysis.

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The compound named "Littoral Zone" included the following keywords: "Sublittoral"; "Intertidal";
"Littoral"; "Beach"; "Dune".

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229 Indicator fauna. To overlay a biotic data layer to the grid defined by morphodynamic state and 230 LAZ, we considered a set of organisms recently highlighted as bioindicators of global relevance 231 (Costa et al., 2020). These latter authors noted that the response to anthropogenic disturbances was 232 related to the species (population, presence) organization level rather than higher (community or 233 assemblage) ones. In this background, no organization level was considered, and single species 234 were considered in the analysis as keywords. Finally, two flagship taxa with high conservation 235 priority were added, such as nesting shorebirds (also mentioned in WIOMSA guidelines) and turtles 236 (see McLachlan et al., 2013). Spawning fish, even though a relevant variable to both ecological and social template, was here not added as they are limited to specific (macrotidal and shallowsublittoral) waters, hence these are not universal.

- The compound term "Fauna" included the following keywords: "Talitrid amphipods"; "Donacid
 clams"; "Ghost crabs"; "Spionid polychaetes"; "Beetles"; "Bird nest"; "Turtle nest".
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- 242 2.2. *Analysis*

243 The keyword co-occurrence analysis was performed following established bibliometric steps of: information retrieval, pattern matching, data analysis, and data visualization (Cobo et al., 2011, 244 245 Callon et al., 1983). A total of 32,304,541 unique abstracts were retrieved from PubMed -246 MEDLINE collection (accessed 03 April 2021). The abstracts were searched for the specific keywords, their synonyms and plural and hyphenated forms. The keyword occurrences in abstracts 247 248 were then transformed to calculate their pairwise co-occurrences (Callon et al., 1983). These co-249 occurrences formed a network, which was analysed and visualised. The code is available 250 here https://github.com/lab42open-team/pubmed trend analysis.

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252 **3. Results**

There has been a large-scale increase of scientific publications targeting plastic (> 80,000 abstracts) and litter (> 20,000 abstracts) (Figure 2). Even though these two keywords have been present in literature since the 1960s, the increase became exponential since the 1990s. Keywords related to plastic sizes, such as "microplastic" and "nanoplastic" appear to be on the same trend, although they started being mentioned in the last two decades. The heatmaps (Figures 2 and 3) show an increase over time in literature as well as the co-dominance of faunal, litter and geomorphological terms.

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Figure 2. Time heatmap of published literature. Keywords are on the Y-axis, grouped in compounds(visualized on the right).

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The keyword "plastic" makes the strongest co-occurrences values: the highest co-occurrence is represented by words "plastic" and "microplastic" (991 abstracts including both words), followed by "plastic" and "litter" (583 abstracts including both words) (Figure 3). These highest cooccurrence values were found within the "litter" compound. There were then 14 pairs of words cooccurring between 100 and 500 times; among them, eight pairs were across different compounds: "turtle" and "beach"; "turtle" and "plastic"; "plastic" and "beach"; "plastic" and "beetle"; "beach" and "litter"; "litter" and "beetle"; "litter" and "fauna"; "intertidal" and "fauna". "Litter" is the compound being mentioned in six of them, in co-occurrence with keywords from the "fauna" and "LAZ" compounds.

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275276 Figure 3. Heatmap of co-occurrences of keywords (in alphabetical order).

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278 The co-occurrence network (Figure 4) highlights the way in which all keywords are generally used, 279 similarly to a random network. The consideration of compounds, however, suggests a clustering, 280 with "microplastic" being more connected to the network than words related to other plastic sizes, 281 which remain at the edges of the network. Also keywords related to morphodynamic state remained 282 at the margin, pointing at a scarce integration in beached plastics litter studies. It has to be noted 283 however, that the lower number of publications (Figure 2) could have played a role in defining this 284 pattern. Finally, the two keywords related to beach-specific life stages of iconic species, i.e. turtle 285 nest and bird nest, remain less connected. In contrast, the trend highlighted for the "LAZ" 286 compound term is revealing that, in spite of the extremely scarce use of the concept (only recovered 287 in the last few years after being neglected for decades), features included in the LAZ are being considered in research - especially "beach" but also "intertidal" and "littoral"- and the concept of 288 289 the active zone could be directly fed with data proceeding from such studies, including those on 290 plastics litter.

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Figure 4. Network visualization of co-occurrences. Compounds are marked in different colours and symbols. Sizes of symbols relate to the degree of co-occurrence of one word with all the others, while the thickness of the line indicates the number of co-occurrences between two single words.

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297 4. Discussion

The very large number of publications targeting plastics appears to include two trends related to macroplastics: 1) while still increasing, publications on macroplastics (unless "litter" and "macroplastic" are used as synonyms) are not increasing as much as those on microplastics, and 2) they remain less related to variables relevant to beach ecology. This latter point might hamper the consideration of a systems approach, where processes are regulated by key ecological variables, necessary to explain and predict the patterns observed. In a dynamic context such as the increasing number of publications on plastics, the detected use of keywords should serve as a warning to scientists, given that published literature – as analysed in this study - is the foundation of prospective research. Integration of newly produced data on plastics litter with known ecologically relevant features of beach ecology should proceed, especially in the view of the UN Decade of Ocean Science for Sustainability 2021-30, which has identified marine litter are a priority topic (Claudet et al., 2020; Elliott, 2021).

Research designed to obtain interoperable and comparable data would allow the ability to fully exploit information, towards advances in both sandy beach ecology and in studies related to beached macroplastics. Considering beaches as social ecological systems (Fanini et al., 2021; Defeo et al., in press) and the tight intertwining of the social and ecological parts of such a system, welldefined in space and easy to identify, such integrated information would promptly find multiple pathways for mainstreaming science evidence into society and governance.

From the analysis were derived potential constraints to integration, which were then grouped into three general topics: 1) unequivocation, 2) identifying a scale for the coupling ecology and plastics on beaches, and 3) targeting dynamics of beached plastics.

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320 4.1. Unequivocation

321 With a rapidly increasing number of publications, and related datasets about plastics, meta-analyses 322 will be required with a clear, unequivocal identification of items. In this respect, the use of terms 323 which have been in use for long, but applied to other disciplines, should be used with caution. As 324 examples, "litter", and "standing stock" are adapted from ecology, although their meaning deeply 325 differs when related to plastics or to natural material. Especially in the case of litter, the ambiguity 326 also extends to a common perception among beach users and the public, i.e. whatever material is 327 found stranded is litter, and is seen as damage to the aesthetics of landscape (see e.g. Williams et 328 al., 2016). The distinction - starting from keywords - of natural vs. plastics substances should 329 remain clear instead, due to their greatly different qualitative effects.

330 Also, terms such as macro-, meso- and micro- imply a range of sizes which differ between plastics and biology studies. The threshold of 5 mm as a discriminating size between macro and micro 331 332 plastics (Frias and Nash 2019, even though this is the most common, is not the only existing 333 definition) does not apply to macro and microorganisms for instance, where the threshold is defined 334 by the ability of resolution by the human eye. However, size categories are essential to unravel the 335 interactions of plastics with beach substratum material and size, with unconsolidated material size 336 spanning from sand classes (63 μ m – 1.5 mm), but also pebbles (2-64 mm) and cobbles (65-512 337 mm) (Blott and Pye, 2001), and a mixture of them. This is especially because of the importance of 338 particle size in defining and interrogating the structure and functioning of beaches.

Recommendation: To avoid issues in current and future information management and analysis, it is recommended to add the word "plastic" whenever it refers to litter and/or standing stock, allowing recognition by improving discoverability in search engines and by text mining tools. The cooccurrence of these terms found in the analysis is a good sign, although such co-mention should become routine. The use of common synonyms (e.g. macroplastic OR macroplastics OR macroplastic OR macro-plastics OR macro plastic OR macro plastics) should also become established for search engines.

Extending terminology standardization, terms such as macro-, meso-, micro- and nano- should consistently be accompanied by a dimension range, by the word "plastic" and used in one form (without a hyphen or space between them).

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351 *4.2.Identifying a scale for coupling ecology and plastics on beaches*

352 A common ground for both beach ecology and beached macroplastics litter studies is the 353 consideration of a single beach as biogeomorphological unit – a mesoscale in sandy beach ecology, 354 where across- and along-shore physical and biotic patterns can be detected. Geomorphological 355 characteristics shape biotic processes on sandy beaches with a well-defined across-shore gradient. It 356 is therefore appropriate to assume that they also shape the interaction of plastics within the system. 357 Broadly used protocols to study macroplastics litter do include some beach variables, although these 358 are not framed in compound terms such as the geomorphological state and the LAZ. This prevents 359 the identification of a beach as a system with boundaries and, as a consequence, the systemic effects 360 of plastics as a stressor. Data on key variables for beach ecology (such as beach width, beach 361 exposure, beach slope, dune presence) are indeed easy to gather, not least from aerial photographs 362 or satellite images, and could frame the analysis of patterns within a systemic vision. In widespread 363 protocols, it is recommended to use standard stretches or areas (e.g. 100 linear m transects as with 364 OSPAR, or standard quadrats). While this allows consistency in the relative presence and 365 abundance of plastics (see e.g. Clean Coast Index, Alkalay et al., 2007 and subsequent index 366 modifications, which are still based on the number of items per area), it does not account for beach key features. Internationally-accepted protocols (Galgani et al, 2013; Hanke et al. 2019; GESAMP, 367 2019; Fleet et al., 2021) also consider freshwater inputs (Riverine Litter Observation Network) and 368 369 urban areas as drivers in marine litter accumulation. However, the selection of units across a 370 gradient of impact is often problematical.

372 Recommendation: By applying the ecological mesoscale (single beach) as the nominated unit, 373 several dimensions for the interaction of beach ecology with beached plastics could be identified. 374 The selection of sites could be less random and include the consideration of the morphodynamic 375 state of beaches (from dissipative to reflective), as well as different substrata, and of the clear 376 identification of the LAZ. In the case of extended beaches, the time/energy cost could be a limiting 377 factor for researchers and/or citizen scientists. In these cases, indications from geomorphology 378 (Nordstrom, 2005) and biodiversity studies (see specifically Battisti et al. 2017, for the application 379 of biodiversity metrics to beached plastics) regarding the selection of subsites and replicates can be 380 useful to optimize resources and create integrated datasets. Essential ecological variables defining 381 the morphodynamic state could be cost-effectively integrated into protocols, given their simple 382 measures: beach width, beach slope, exposure, grain size, and/or salinity. Furthermore, by 383 considering single beaches as the unit for research across gradients, the concept of the gravity centre 384 (Peng et al., 2017) could be developed to highlight spatial patterns such as those defined by cities 385 and main freshwater discharges, and also to indicate temporal patterns (e.g. seasonal use of the 386 beach). Finally, patterns related to a relevant ecological dimension could be connected to the social 387 one, providing insights of a shift from reactive studies to proactive ones (Cinner, 2018).

388 4.3. Targeting dynamics of beached plastics

389 To address the problem of plastic pollution, it is of paramount importance to interrogate patterns 390 observed with system drivers and dynamics, enabling the formulation of strategies and actions. 391 Once the boundaries of the system are identified, the classification of internal and external drivers 392 will follow logically, placing the information (which might be already largely available from 393 existing datasets) as tiles in a mosaic. The LAZ was proposed as the unit relevant at the ecological 394 and social-economic levels for the depiction of dynamics connecting these two states (Fanini et al., 395 2021) and could be considered as a unit also in the case of beached plastic studies. For example, 396 hydrological or meteorological drivers, which may be important for budgeting or analysing 397 dynamics of macroplastics on beaches, would act on the LAZ. Similarly, social drivers are also 398 acting on the LAZ. In this respect, some good practices are already routinely established, such as 399 the brand audits on beached items (for example using the bar-coding on labels), allowing the 400 identification of dynamics of contamination and pollution (e.g. the age and source of the plastics). 401 Many other actions at different scales might be explored to analyse the dynamics connecting 402 producers/users/actors in charge of disposal, matching them with the patterns observed and reported 403 in publications. Actions finely tuned to the specific context could be proposed, targeting, for 404 example, the reduction of use and alternative choices to plastics (Riechers et al., 2021), as well as

405 monitoring tools. Some of the LAZ components were found linked to beached plastics litter studies 406 and so a data background is likely to be readily available following the conceptual up-take of the 407 LAZ as part of a systems analysis. Temporal dynamics also deserve attention given that 408 microplastics, as the degradation products of plastics and litter, have lately received a large amount 409 of attention (Ivleva et al., 2017; Ryan et al., 2015; Rodrigues et al., 2021). Hence, it is timely to 410 discriminate between primary and secondary particles, which is the dynamic connection between 411 macro- and microplastic. Tools are increasingly available for the identification of plastic material 412 found stranded, supporting essential information, such as toxicity, inertia, weathering (including 413 biofilm creation) and break-down likelihood and follow up paths related to the occurrence of 414 primary and secondary particles of plastics on a beach (Rodrgues et al., 2021). A focus on the 415 weathering and breakdown of items on beaches might be a suitable inference method to link to 416 studies on beach dynamics which started more than 50 years ago (see e.g. Frigione et al., 2021).

417 Recommendations: As with the budgetary approach to the dynamics of soft substratum – a concept 418 on which the LAZ was originally based - budgetary approaches can also be established for plastics. 419 Inputs and outputs into the LAZ could be estimated over different temporal scales, but also in terms 420 of macro- and micro-plastic fractions (in terms of both weight and number of items, as already suggested in international protocols). This will shed further light on the eventual inter-dependency 421 422 of sizes, especially if paired with the identification of social (e.g. tourism; fishery) and natural 423 ecological/environmental (e.g. monsoons, beach exposure) drivers. Studies discriminating between 424 primary and secondary microplastics should be encouraged, as they would greatly support the 425 understanding of breakdown dynamics of plastics (GESAMP, 2015) while beached.

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427 5. Conclusions

428 As remarked by Borja and Elliott (2019), it is no longer time to report occurrences of plastics 429 without proposing solutions. It is also timely to tailor general solutions such as "increasing 430 awareness; reducing littering; etc." to the specific context, i.e. defining system components, 431 boundaries, and dynamics of interaction. Available data would then fit into such a systematic 432 vision, allowing the elucidation of paths, on which calibrated solutions can be proposed and hold a 433 higher likelihood of success. However, published literature showed that the coupling between 434 plastic studies and the geomorphological beach system (the very background of its definition) is still 435 limited. Therefore, the huge potential arising from integrated data collection still needs to be 436 revealed. Integration could ultimately support governance, enhancing the return of research results 437 as policy-informing and operational knowledge, especially in the case of beached plastics litter. 438 This would counter the current trend in which beach managers and stakeholders are only exposed to

a one-size-fits-all regulation with respect to beached plastics, whatever the exposure of the beach to waves and tides, and the size of the substratum particles. The consideration and inclusion of local characteristics would greatly sustain the small-scale management, often neglected by national and international guidelines. If intrinsic beach characteristics remain disconnected from monitoring programmes and we do not capitalize on the information available from beach ecology, there is the high risk of not increasing our understanding thereby disconnecting macroplastics litter studies from those beach features defining functional stability and ultimately, environmental sustainability.

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- 453 References

Addamo, A. M., Laroche, P., Hanke, G. 2017. Top Marine Beach Litter Items in Europe, EUR
29249 EN, Publications Office of the European Union, Luxembourg, doi:10.2760/496717,
JRC108181

457

Alkalay, R., Pasternak, G., Zask, A. 2007. Clean-coast index—a new approach for beach
cleanliness assessment. Ocean Coast. Manag., 50, 352-362.
https://doi.org/10.1016/j.ocecoaman.2006.10.002

461

462 AmesWeb, <u>https://amesweb.info/Materials/Density-of-Plastics.aspx</u>

463 [Accessed 10/05/2021]

464

Andrady, A. L. 2011. Microplastics in the marine environment. Mar. Pollut. Bull., 62, 1596-1605.
https://doi.org/10.1016/j.marpolbul.2011.05.030

- 467
- 468

Barnardo, T., Ribbink, A.J. (Eds.). 2020. African Marine Litter Monitoring Manual. African Marine
Waste Network, Sustainable Seas Trust. Port Elizabeth, South Africa.
http://dx.doi.org/10.25607/OBP-923

- Battisti, C., Bazzichetto, M., Poeta, G., Pietrelli, L., Acosta, A.T., 2017. Measuring non-biological
 diversity using commonly used metrics: Strengths, weaknesses and caveats for their application in
 beach litter management. J Coast. Conserv., 21, 303-310. doi 10.1007/s11852-017-0505-9
 Blott, S.J., Pye, K. 2001. GRADISTAT: a grain size distribution and statistics package for the
- analysis of unconsolidated sediments. Earth surface processes and Landforms, 26, 1237-1248.
 <u>https://doi.org/10.1002/esp.261</u>
- 480
- 481
- Borja, A. Elliott, M. 2019. So when will we have enough papers on microplastics and ocean litter?
 Mar. Pollut. Bull., 146, 312-316. http://dx.doi.org/10.1016/j.marpolbul.2019.05.069
- 484
- 485 Callon, M., Courtial, J.-P., Turner, W.A., Bauin, S. 1983. From translations to problematic
 486 networks: An introduction to co-word analysis. Soc. Sci. Inform., 22, 191–235.
 487 <u>https://doi.org/10.1177/053901883022002003</u>
- 488

Cheshire, A.C., Adler, E., Barbière, J., Cohen, Y., Evans, S., Jarayabhand, S., Jeftic, L., Jung, R.T.,
Kinsey, S., Kusui, E.T., Lavine, I., Manyara, P., Oosterbaan, L., Pereira, M.A., Sheavly, S., Tkalin,
A., Varadarajan, S., Wenneker, B., Westphalen, G. 2009. UNEP/IOC Guidelines on Survey and
Monitoring of Marine Litter. UNEP Regional Seas Reports and Studies, No. 186; IOC Technical
Series No. 83: xii + 120 pp.

- 494
- Cinner, J.E., Adger, W.N., Allison, E.H., Barnes, M.L., Brown, K., Cohen, P.J., Gelcich, S., Hicks,
 C.C., Hughes, T.P., Lau, J. and Marshall, N.A. 2018. Building adaptive capacity to climate change
 in tropical coastal communities. Nat. Clim. Change, 8, 117-123.
- 498
- Claudet, J., Bopp, L., Cheung, W.W., Devillers, R., Escobar-Briones, E., Haugan, P., Heymans, J.J.,
 Masson-Delmotte, V., Matz-Lück, N., Miloslavich, P. and Mullineaux, L. 2020. A roadmap for
 using the UN decade of ocean science for sustainable development in support of science, policy,
 and action. One Earth, 2, 34-42. <u>https://doi.org/10.1016/j.oneear.2019.10.012</u>
- 503
- Cobo, M.J., López-Herrera, A.G., Herrera-Viedma, E., Herrera, F. 2011. Science mapping software
 tools: Review, analysis, and cooperative study among tools. J Am Soc Inf Sci Tec 62, 1382–1402.
 https://doi.org/10.1002/asi.21525

507	
508	Cormier, R., Elliott, M. 2017. SMART marine goals, targets and management - is SDG 14
509	operational or aspirational, is 'Life Below Water' sinking or swimming? Mar. Pollut. Bull., 123, 28-
510	33. https://doi.org/10.1016/j.marpolbul.2017.07.060
511	
512	Costa, L. L., Zalmon, I. R., Fanini, L., Defeo, O. (2020). Macroinvertebrates as indicators of
513	human disturbances on sandy beaches: A global review. Ecol. Indic. 118, 106764.
514	https://doi.org/10.1016/j.ecolind.2020.106764
515	
516	
517	Cresta, E., Battisti, C. 2021. Anthropogenic litter along a coastal-wetland gradient: Reed-bed
518	vegetation in the backdunes may act as a sink for expanded polystyrene. Mar. Pollut. Bull., 172,
519	112829. https://doi.org/10.1016/j.marpolbul.2021.112829
520	
521	
522	Defeo, O, Elliott, M. 2021. The 'Triple Whammy' of coasts under threat - why we should be
523	worried! Mar. Pollut. Bull., 163, 111832 Doi 10.1016/j.marpolbul.2020.111832
524	
525	Defeo, O., McLachlan, A., Armitage, D., Elliott, M., Pittman, J. (in press). Sandy beach social-
526	ecological systems at risk: regime shifts, collapses and governance challenges. Frontiers in Ecology
527	and the Environment Doi: 10.1002/fee.2406.
528	
529	Elliott, M. 2021. Marine pollutants and contaminants: marine problems, solutions and the role of
530	the UN Decade of Ocean Science for Sustainable Development. Environmental SCIENTIST
531	(Journal of the Institution of Environmental Scientists) 30, 10-17.
532	
533	Fanini, L., Defeo, O., Elliott, M., 2020. Advances in sandy beach research-Local and global
534	perspectives. Estuar. Coast. Shelf S., 234:106646.
535	
536	Fanini, L., Piscart, C., Pranzini, E., Kerbiriou, C., Le Viol, I., Pétillon, J., 2021. The extended
537	concept of littoral active zone considering soft sediment shores as social-ecological systems, and an
538	application to Brittany (North-Western France). Estuar. Coast. Shelf S., 250:107148.
539	

540	Fleet, D., Vlachogianni, T,. Hanke, G. 2021. A Joint List of Litter Categories for Marine Macrolitter
541	Monitoring. EUR 30348 EN, Publications Office of the European Union, Luxembourg, 2021,
542	doi:10.2760/127473, JRC121708
543	
544	Frias, J.P.G.L., Nash, R. 2019. Microplastics: finding a consensus on the definition. Mar. Pollut.
545	Bull., 138,145-147.
546	
547	Frigione, M., Marini, G., Pinna, M. 2021. A Thermal Analysis-Based Approach to Identify
548	Different Waste Macroplastics in Beach Litter: The Case Study of Aquatina di Frigole NATURA
549	2000 Site (IT9150003, Italy). Sustainability, 13, 3186.
550	
551	Galgani, F., Hanke, G., Werner, S., Oosterbaan, L., Nilsson, P., Fleet, D., Kinsey, S., RC, T., Van
552	Franeker, J., Vlachogianni, T., Scoullos, M., Mira Veiga, J., Palatinus, A., Matiddi, M., Maes, T.,
553	Korpinen, S., Budziak, A., Leslie, H., Gago, J., Liebezeit, G. 2013. Guidance on Monitoring of
554	Marine Litter in European Seas. European Commission, Joint Research Centre (2013). MSFD
555	Technical Subgroup on Marine Litter (TG ML). EUR 26113; doi: 10.2788/99475.
556	
557	
558	GESAMP, 2015. Sources, fate and effects of microplastics in the marine environment: a global
559	assessment (Kershaw, P.J. ed.), Rep. Stud. GESAMP No. 90, 96 p.
560	
561	GESAMP, 2019. Guidelines or the monitoring and assessment of plastic litter and microplastics in
562	the ocean (Kershaw P.J., Turra A. and Galgani F. editors), (IMO/FAO/UNESCO-
563	IOC/UNIDO/WMO/IAEA/UN/UNEP/UNDP/ISA Joint Group of Experts on the Scientific Aspects
564	of Marine Environmental Protection). Rep. Stud. GESAMP No. 99, 130 p.
565	
566	Hanke G., Walvoort D., van Loon W., Addamo A.M., Brosich A., del Mar Chaves Montero M.,
567	Molina Jack M.E., Vinci M., Giorgetti A. 2019. EU Marine Beach Litter Baselines, EUR 30022
568	EN, Publications Office of the European Union, Luxemburg. doi:10.2760/16903, JRC114129.
569	
570	Haklay M., Dörler D., Heigl F., Manzoni M., Hecker S., Vohland K. 2021. What Is Citizen
571	Science? The Challenges of Definition. In: Vohland K. et al. (eds) The Science of Citizen Science.
572	Springer, Cham. https://doi.org/10.1007/978-3-030-58278-4_2
573	

574	HELCOM 2008. HELCOM Recommendation 29/2. Marine litter within the Baltic Sea region. Last
575	accessed 13 June 2013 online at: http://www.helcom.fi/Recommendations/en_GB/rec29_2/
576	
577	Hidalgo-Ruz, V., Thiel, M. 2015. The Contribution of Citizen Scientists to the Monitoring of Marine
578	Litter. In: Bergmann, M., Gutow, L. and Klages, M (eds.). Marine anthropogenic litter, . Springer
579	Nature, 447 p.
580	
581	Ivleva, N. P., Wiesheu, A. C., Niessner, R. 2017. Microplastic in aquatic ecosystems. Angew.
582	Chem. Int. Edit., 56, 1720-1739. https://doi.org/10.1002/anie.201606957
583	
584	JRC scientific and policy reports, 2013. Guidance on monitoring of litter in European Marine Seas",
585	MSFD Technical Subgroup on Marine Litter, , doi:10.2788/99475
586	
587	Lambert S., Sinclair C., Boxall A. (2014) Occurrence, Degradation, and Effect of Polymer-Based
588	Materials in the Environment. In: Whitacre D. (eds) Reviews of Environmental Contamination and
589	Toxicology, Volume 227. Reviews of Environmental Contamination and Toxicology (Continuation
590	of Residue Reviews), vol 227. Springer, Cham. https://doi.org/10.1007/978-3-319-01327-5_1
591	Lebreton, L., Andrady, A. 2019. Future scenarios of global plastic waste generation and disposal.
592	Palgrave Communications, 5, 6. https://doi.org/10.1057/s41599-018-0212-7
593	
594	McLachlan, A., 1983. Sandy beaches as ecosystems., In: Sandy beach ecology-a review. Springer
595	pp.321-380. Doi: 10.1007/978-94-017-2938-3_25
596	
597	McLachlan, A., Defeo, O., Jaramillo, E., Short, A.D. 2013. Sandy beach conservation and
598	recreation: guidelines for optimising management strategies for multi-purpose use. Ocean Coast.
599	Manage., 71, 256-268. https://doi.org/10.1016/j.ocecoaman.2012.10.005
600	
601	McLachlan, A. and Defeo, O., 2018. The ecology of sandy shores. Academic Press.
602	
603	McLachlan, A., Defeo, O., Short, A.D., 2018. Characterising sandy beaches into major types and
604	states: Implications for ecologists and managers. Estuar. Coast. Shelf S., 215, 152-160.
605	

- Morales-Caselles, C., Viejo, J., Martí, E. *et al.* 2021. An inshore–offshore sorting system revealed
 from global classification of ocean litter. Nat. Sustain. 4, 484–493. https://doi.org/10.1038/s41893021-00720-8
- Moreira, F.T., Balthazar-Silva, D., Barbosa, L., Turra, A., 2016. Revealing accumulation zones of
 plastic pellets in sandy beaches. Environ. Pollut. 218, 313-321.
 <u>https://doi.org/10.1016/j.envpol.2016.07.006</u>
- 612
- 613 Neff, M.W., Corley, E.A. 2009. 35 years and 160,000 articles: A bibliometric exploration of the
- 614 evolution of ecology. Scientometrics 80, 657–682. <u>https://doi.org/10.1007/s11192-008-2099-3</u>
- 615
- Nordstrom, K. F. 2005. Beach nourishment and coastal habitats: research needs to improve
 compatibility. Restor. Ecol., 13, 215-222. <u>https://doi.org/10.1111/j.1526-100X.2005.00026.x</u>
- 618
- Opfer, S., Arthur, C., Lippiatt, S. 2012. NOAA Marine Debris Shoreline Survey Field Guide.
 National Oceanic and Atmospheric Administration http://dx.doi.org/10.25607/OBP-937
- 621
- 622 OSPAR Commission 2010. Wenneker, B.; Oosterbaan, L. and Intersessional Correspondence 623 Group on Marine Litter (ICGML) (2010) Guideline for Monitoring Marine Litter on the Beaches in 624 Maritime Area. Edition 1.0. the OSPAR London, UK, 15pp. & Annexes. DOI: 625 http://dx.doi.org/10.25607/OBP-968
- 626
- Peng, J., Zhao, M., Guo, X., Pan, Y., Liu, Y. 2017. Spatial-temporal dynamics and associated
 driving forces of urban ecological land: A case study in Shenzhen City, China. Habitat Int., 60,8190. <u>https://doi.org/10.1016/j.habitatint.2016.12.005</u>
- 630
- Rangel-Buitrago, N., Williams, A., Anfuso, G., Arias, M., Gracia, A. 2017. Magnitudes, sources,
 and management of beach litter along the Atlantico department coastline, Caribbean coast of
 Colombia. Ocean Coast. Manag. 138, 142-157. <u>https://doi.org/10.1016/j.ocecoaman.2017.01.021</u>
- 634
- Rees, G., Pond, K. 1995. Marine litter monitoring programmes -A review of methods with special
 reference to national surveys. Mar. Pollut. Bull. 30, 103 -108. <u>https://doi.org/10.1016/0025-</u>
 <u>326X(94)00192-C</u>
- 638

639	Riechers, M., Fanini, L., Apicella, A., Galván, C.B., Blondel, E., Espiña, B., Kefer, S., Kere	oullé, T.,				
640	Klun, K., Pereira, T.R., Ronchi, F., 2021. Plastics in our ocean as transdisciplinary challenge. Mar.					
641	Pollut. Bull. 164,112051. https://doi.org/10.1016/j.marpolbul.2021.112051					
642						
643	Riverine Litter Observation	Network,				
644	https://mcc.jrc.ec.europa.eu/main/dev.py?N=simple&O=394&titre_page=RIMMEL%2520c	observati				
645	on%2520Network [Last accessed 14 May 2021]					
646						
647	Rodrigues, S.M., Elliott, M., Almeida, C.M.R., Ramos, S.2021. Microplastics and p	plankton:				
648	knowledge from laboratory and field studies to distinguish contamination from poll	ution. J.				
649	Hazard. Mater., 417, 126057, https://doi.org/10.1016/j.jhazmat.2021.126057.					
650						
651	Ryan, P. G. 2015. A brief history of marine litter research. In: Marine anthropogenic litte	er (pp. 1-				
652	25). Springer, Cham. https://doi.org/10.1007/978-3-319-16510-3_1					
653						
654	Ryan, P.G., Perold, V. 2021. Limited dispersal of riverine litter onto nearby beaches during	g rainfall				
655	events. Estuar. Coast. Shelf Sci., 251,107186.					
656						
657						
658	Schulz, M., van Loon, W., Fleet, D.M., Baggelaar, P., van der Meulen, E., 2017. OSPAR	standard				
659	method and software for statistical analysis of beach litter data. Mar. Pollut. Bull., 122:	166-175.				
660	https://doi.org/10.1016/j.marpolbul.2017.06.045					
661						
662	Syberg, K., Palmqvist, A., Khan, F.R., Strand, J., Vollertsen, J., Clausen, L.P.W., I	Feld, L.,				
663	Hartmann, N.B., Oturai, N., Møller, S., Nielsen, T.G., 2020. A nationwide assessment of	of plastic				
664	pollution in the Danish realm using citizen science. Sci. Rep., 10,	, 1-11.				
665	https://doi.org/10.1038/s41598-020-74768-5					
666						
667	Tinley, K. L. 1985. Coastal Dunes of South Africa 109FRD, CSIR, Pretoria, South Africa, 3	300 pp.				

669 Turra, A., Manzano, A.B., Dias, R.J.S., Mahiques, M.M., Barbosa, L., Balthazar-Silva, D., Moreira,

670 F.T. 2014. Three-dimensional distribution of plastic pellets in sandy beaches: shifting paradigms.

671 Sci. Rep., 4,1-7. Doi: 10.1038/srep04435

- 672
- UN, 2015 Transforming our world: the 2030 agenda for sustainable development Resolution
 Adopted by the General Assembly on 25 September 2015. Seventieth Session, Agenda Items 15
 and 116. A/RES/70/1. <u>https://www.un.org/ga/search/view_doc.asp?symbol=A/RES/70/1&Lang=E</u>
- UNEP Regional Seas Programme, UNEP. Mediterranean Action Plan, Secretariat of the Basel
 Convention on the Control of Transboundary Movements of Hazardous Wastes, Their Disposal,
 UNEP/GPA Coordination Office, & Intergovernmental Oceanographic Commission. 2005. Marine
 Litter: An Analytical Overview
- 681
- Van Loon, W., Hanke, G., Fleet, D., Werner, S., Barry, J., Strand, J., Eriksson, J., Galgani, F.,
 Gräwe, D., Schulz, M., Vlachogianni, T., Press, M., Blidberg, E., Walvoort, D., 2020. A European
 Threshold Value and Assessment Method for Macro Litter on Coastlines. EUR 30347 EN,
 Publications Office of the European Union, Luxembourg, 2020, doi:10.2760/54369, JRC121707
- 686
- Williams, A.T., Tudor, D.T., 2001. Litter burial and exhumation: spatial and temporal distributionon a cobble pocket beach. Mar. Pollut. Bull., 42, 1031-1039.
- 689
- Williams, A.T., Rangel-Buitrago, N.G., Anfuso, G., Cervantes, O., Botero, C.M., 2016. Litter
 impacts on scenery and tourism on the Colombian north Caribbean coast. Tourism Manage., 55,
 209-224.







